Terrestrial nitrogen–carbon cycle interactions at the global scale

S. Zaehle

Department of Biogeochemical Integration, Max Planck Institute for Biogeochemistry, Hans-Knoell-Strasse 10, 07745 Jena, Germany

Interactions between the terrestrial nitrogen (N) and carbon (C) cycles shape the response of ecosystems to global change. However, the global distribution of nitrogen availability and its importance in global biogeochemistry and biogeochemical interactions with the climate system remain uncertain. Based on projections of a terrestrial biosphere model scaling ecological understanding of nitrogen–carbon cycle interactions to global scales, anthropogenic nitrogen additions since 1860 are estimated to have enriched the terrestrial biosphere by 1.3 Pg N, supporting the sequestration of 11.2 Pg C. Over the same time period, CO₂ fertilization has increased terrestrial carbon storage by 134.0 Pg C, increasing the terrestrial nitrogen stock by 1.2 Pg N. In 2001–2010, terrestrial ecosystems sequestered an estimated total of 27 Tg N yr⁻¹ (1.9 Pg C yr⁻¹), of which 10 Tg N yr⁻¹ (0.2 Pg C yr⁻¹) are due to anthropogenic nitrogen deposition. Nitrogen availability already limits terrestrial carbon sequestration in the boreal and temperate zone, and will constrain future carbon sequestration in response to CO₂ fertilization (regionally by up to 70% compared with an estimate without considering nitrogen–carbon interactions). This reduced terrestrial carbon uptake will probably dominate the role of the terrestrial nitrogen cycle in the climate system, as it accelerates the accumulation of anthropogenic CO₂ in the atmosphere. However, increases of N₂O emissions owing to anthropogenic nitrogen and climate change (at a rate of approx. 0.5 Tg N yr⁻¹ per 1°C degree climate warming) will add an important long-term climate forcing.

1. Introduction

Nitrogen is a fundamental component of living organisms. Ecosystem available forms of nitrogen (ammonium, as well as nitrate among other oxidized nitrogen forms), hereafter reactive N (Nᵣ), are scarce in unperturbed ecosystems owing to low atmospheric inputs, the high energetic costs of assimilating elementary N₂ through biological fixation and nitrogen losses to leaching and volatilization, particularly after disturbances [1]. The productivity of plants and soil organisms strongly depends on nitrogen, imposing stoichiometric constraints at the level of an individual organism. These two facts lead to a tight coupling of the terrestrial nitrogen and carbon cycles, as evidenced by the constrained flexibility of ecosystem C:N stoichiometry [2]. N availability thereby plays an important role in controlling the productivity, structure and spatio-temporal dynamics of terrestrial ecosystems: perturbations in the nitrogen cycle will have repercussions in the carbon cycle, and vice versa.

The terrestrial biogeochemical cycles have been disturbed in the past by human actions altering land cover and land-use, by increasing the atmospheric abundance of CO₂, and by doubling the inputs of Nᵣ through the burning of fossil fuel and the creation of agricultural fertilizer since 1860 [3,4]. These anthropogenic changes must have had consequences for the terrestrial store and turnover of nitrogen and carbon. However, because of the uncertainty in (i) the global distribution of nitrogen availability and demand in terrestrial ecosystems, (ii) the capacity of the terrestrial biosphere to retain added nitrogen and (iii) the tightness of the coupling between the terrestrial nitrogen and carbon cycles, these consequences are not well understood. The regional distribution of the anthropogenic perturbation is also important to take into account, as the fertilization by anthropogenic CO₂—even if regionally constrained by nitrogen availability—is ubiquitous, whereas high
levels of anthropogenic N, only affect a small fraction of the global land surface, and land-use changes mainly act locally.

Quantifying the changes in the terrestrial carbon and nitrogen budgets is relevant not only to understand the fate of the anthropogenic N, and the cascading effects of this nitrogen, but also because these changes matter for the climate system [4]. Limited natural N availability reduces the carbon storage potential of the terrestrial biosphere. Anthropogenic N deposition generally increases terrestrial C sequestration and thus decreases the rate of anthropogenic CO₂ accumulation in the atmosphere, but at the same time enhances nitrogen losses for instance to the greenhouse gas N₂O, which might compensate for the C-cycle-related climate benefit [5,6]. This is important because the long atmospheric lifetime of the N₂O can transform even subtle but long-term changes in terrestrial emission into a significant climate forcing.

The objective of this paper is to provide an assessment of the present and future nitrogen–carbon cycle interactions with a focus on the role of the natural and perturbed terrestrial nitrogen cycles in shaping the terrestrial net carbon and nitrogen balance and terrestrial carbon–climate feedbacks. A suite of new global ecosystem models that integrates current ecological and biogeochemical understanding with process-based descriptions of the terrestrial energy and water balance at a comparatively high spatial resolution is now available for such a task [7]. However, no systematic and comprehensive analyses have been performed so far with several models that would allow for a systematic model synthesis. I therefore present past, present and future nitrogen and carbon budgets based on one model only, the O–CN model [6,8], and discuss the uncertainties related to the application of this model in the light of other modelling studies and independent estimates.

2. Material and methods

(a) The O–CN model

O–CN [6,8] is a terrestrial biosphere model, which has been developed from the land surface model ORCHIDEE [9], and describes the nitrogen and carbon fluxes and stocks of vegetation and soil organic matter for 10 natural plant functional types, as well as C3 and C4 croplands at a half hourly time scale. The biogeochemical fluxes are tightly coupled to the calculations of the terrestrial energy and water balance. Nitrogen availability directly controls photosynthesis and respiration of vegetation through tissue nitrogen concentrations and effects on plant allocation (e.g. the root : shoot ratio), and thus foliage area and root growth. Nitrogen availability also affects the temperature-sensitive rate of organic matter decomposition and the net mineralization of nitrogen. The stoichiometry of plant tissues, litter and soil organic matter varies prognostically within observed limits, depending on the relative availabilities of nitrogen and carbon. The modelled ecosystem receives N inputs from biological nitrogen fixation and atmospheric N deposition, and simulates losses of nitrogen to leaching and volatilization based on the process-based simulation of nitrification and denitrification. Fertilizer is applied to the crop-land fraction of each model grid cell at distinct dates during the growing season, but the treatment of cropland management and biomass removal is very simple, and manure systems are not taken into account. O–CN does not simulate the industrial sources and atmospheric transport of N. The model has been evaluated and applied to study the interactions of nitrogen and carbon cycling over the past few decades, and was found to simulate carbon and nitrogen fluxes that are generally commensurate with current understanding [6,8,10,11].

(b) Modelling protocol

O–CN was applied at a 3.75° × 2.5° spatial resolution. The model was brought into steady-state for 1860 conditions and then run transiently in a factorial design to identify the contribution of the individual driving forces. To isolate the effects of nitrogen dynamics, the model has been run twice, once with explicit accounting for nitrogen dynamics (referred to as O–CN), and once with nitrogen concentrations set to global averages of observed values (referred to as O–C), such that plant productivity and soil organic matter decomposition correspond to an ecosystem with average nitrogen availability not taking account of the spatial–temporal patterns of N availability. Two sets of simulations were performed: a ‘historic’ run driven by observed or reconstructed changes in land use, climate, atmospheric CO₂, cropland fertilization and atmospheric deposition (1860–2010), and a ‘future’ run (1860–2100) with a reduced set of forcings (climate, atmospheric CO₂ and N deposition) for the Intergovernmental Panel on Climate Change (IPCC) Special Report on Emissions Scenarios (SRES) A2 scenario.

(c) Datasets

(i) Historic run

The climate forcing (1901–2010) was taken from the CRU-NCEP dataset [4, 12] in the spatially degraded form (at 3.75° × 2.5° spatial resolution) provided by C. Huntingford (2012, personal communication). Gridded time series of cropland fertilization rates [6] and annual land-use changes [13] were used for the period 1860–2005, and assumed constant thereafter. Biological nitrogen fixation in natural vegetation was prescribed based on a climatology developed after Cleveland et al. [11,14]. To assess the uncertainty related to the estimation of nitrogen deposition, decadal time slices of monthly nitrogen deposition fields were obtained from two atmospheric chemistry transport models (CTMs), TM5 [15] and NCAR-CTM [16], and linearly interpolated to arrive at annual values. No estimates beyond 2000 were available for TM5. These were constructed by extrapolating the TM5 estimate for 2000 to 2001–2010 using the grid-cell wise monthly trends of the NCAR-CTM.

(ii) Future run

The future projections are described in detail by Zaehle et al. [10]. The simulations were forced with the SRES A2 climate and atmospheric CO₂ change scenario of the IPSL-CM4 climate model [17], and a nitrogen deposition scenario, which increases the deposition on land from 10 Tg N yr⁻¹ (1860) to 51 Tg N yr⁻¹ (1993) and 106 Tg N yr⁻¹ (2050), after which it was assumed to be constant [3]. This roughly follows the upper boundary of the representative concentration pathway (RCP) nitrogen deposition scenarios [18].

3. Results

(a) Current global terrestrial nitrogen and carbon budget

The contemporary nitrogen and carbon budget displayed in figure 1 (see also table 1) for 2001–2010 is based on the ‘unperturbed’ state of the cycles (1860s) and the historic changes in land cover, climate, atmospheric CO₂ abundance and anthropogenic N inputs from atmospheric deposition and fertilizer application between the years 1860 and 2010.

(i) Nitrogen budget

During 2001–2010, direct (N₅ additions) or indirect (land-use change, climate change and increase in atmospheric CO₂) anthropogenic factors are responsible for 0.5 Pg N of the nitrogen stored in vegetation (15% of the global total), and for
fertilizer: 99.7 (+0 + 85.7)

atmospheric deposition: 61.9 (+0 + 42.1)

BNF: 98.7 (−6.8 − 2.6)

N2: 105.8 (+0.1 + 44.7)

N2O: 8.7 (+0.5 + 2.3)

NOx: 8.5 (+0.7 + 1.3)

human appropriation: 15.0 (+7.3 + 2.1)

leaching loss: 97.1 (+3.2 + 31.2)

Table 1. Global and continental carbon and nitrogen budgets for the years 2001–2010 derived from the O–CN simulations driven with NCAR (TSM) nitrogen deposition fields. BNF, biological nitrogen fixation; GPP, gross primary production; NBP, net biome production = net ecosystem production — anthropogenic C losses.

<table>
<thead>
<tr>
<th>Area (10^6 km^2)</th>
<th>Africa</th>
<th>Asia</th>
<th>Europe</th>
<th>N. America</th>
<th>S. America</th>
<th>Oceania</th>
<th>Global</th>
</tr>
</thead>
<tbody>
<tr>
<td>Nitrogen inputs (Tg N yr^-1)</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>BNF</td>
<td>26.9</td>
<td>20.5</td>
<td>4.3</td>
<td>11.6</td>
<td>29.9</td>
<td>5.2</td>
<td>98.7 (99.1)</td>
</tr>
<tr>
<td>deposition</td>
<td>9.7</td>
<td>27.4</td>
<td>7.5</td>
<td>9.1</td>
<td>5.9</td>
<td>1.2</td>
<td>61.9 (55.6)</td>
</tr>
<tr>
<td>fertilizer</td>
<td>7.2</td>
<td>53.9</td>
<td>13.6</td>
<td>17.9</td>
<td>4.9</td>
<td>1.4</td>
<td>99.7 (99.7)</td>
</tr>
<tr>
<td>Nitrogen losses (Tg N yr^-1)</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>leaching loss</td>
<td>16.1</td>
<td>37.2</td>
<td>7.2</td>
<td>14.2</td>
<td>17.4</td>
<td>3.6</td>
<td>97.1 (92.8)</td>
</tr>
<tr>
<td>volatilization (NH3 + NOx + N2O + N2)</td>
<td>26.0</td>
<td>47.0</td>
<td>12.4</td>
<td>16.9</td>
<td>17.8</td>
<td>4.4</td>
<td>124.9 (123.8)</td>
</tr>
<tr>
<td>soil N2O emission</td>
<td>2.3</td>
<td>2.8</td>
<td>0.5</td>
<td>1.0</td>
<td>1.7</td>
<td>0.4</td>
<td>8.7 (8.7)</td>
</tr>
<tr>
<td>soil N2 emission</td>
<td>20.7</td>
<td>41.4</td>
<td>10.9</td>
<td>15.0</td>
<td>14.0</td>
<td>3.6</td>
<td>105.8 (104.9)</td>
</tr>
<tr>
<td>N export</td>
<td>2.5</td>
<td>5.4</td>
<td>2.8</td>
<td>2.3</td>
<td>1.8</td>
<td>0.2</td>
<td>15.0 (15.0)</td>
</tr>
<tr>
<td>Nitrogen balance (Tg N yr^-1)</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>plant uptake</td>
<td>161.5</td>
<td>205.9</td>
<td>95.5</td>
<td>113.8</td>
<td>201.1</td>
<td>21.4</td>
<td>800.1 (792.2)</td>
</tr>
<tr>
<td>net ecosystem N storage</td>
<td>0.6</td>
<td>13.3</td>
<td>3.3</td>
<td>5.6</td>
<td>4.0</td>
<td>0.1</td>
<td>27.1 (26.8)</td>
</tr>
<tr>
<td>Nitrogen pools (Pg N)</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>vegetation N</td>
<td>0.7</td>
<td>0.9</td>
<td>0.3</td>
<td>0.4</td>
<td>1.1</td>
<td>0.1</td>
<td>3.5 (3.4)</td>
</tr>
<tr>
<td>litter and soil N</td>
<td>13.2</td>
<td>35.7</td>
<td>16.8</td>
<td>21.7</td>
<td>13.7</td>
<td>2.6</td>
<td>103.9 (102.2)</td>
</tr>
<tr>
<td>soil mineral N (10^-3)</td>
<td>14.8</td>
<td>35.5</td>
<td>10.8</td>
<td>15.8</td>
<td>9.8</td>
<td>3.1</td>
<td>90.8 (85.5)</td>
</tr>
<tr>
<td>Carbon fluxes (Pg C yr^-1)</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>GPP</td>
<td>25.6</td>
<td>29.9</td>
<td>11.4</td>
<td>16.8</td>
<td>34.9</td>
<td>3.6</td>
<td>122.3 (121.6)</td>
</tr>
<tr>
<td>NBP</td>
<td>0.1</td>
<td>0.4</td>
<td>0.4</td>
<td>0.5</td>
<td>0.5</td>
<td>0.1</td>
<td>1.9 (1.9)</td>
</tr>
<tr>
<td>Carbon pools (Pg C)</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>vegetation C</td>
<td>114.5</td>
<td>145.8</td>
<td>42.1</td>
<td>75.8</td>
<td>187.3</td>
<td>12.4</td>
<td>581.4 (578.1)</td>
</tr>
<tr>
<td>litter and soil C</td>
<td>169.2</td>
<td>431.2</td>
<td>196.6</td>
<td>270.3</td>
<td>174.5</td>
<td>37.5</td>
<td>1287.6 (1268.2)</td>
</tr>
</tbody>
</table>
1.6 Pg N (2%) stored in soils and litter (excluding wetlands and permafrost soils; figure 2). The most significant cause for vegetation N changes prior to the 1960s has been forest clearance, which has only partly been compensated by increased sequestration owing to CO₂ fertilization. Anthropogenic Nr plays an increasing role after 1960, but remains only a modest cause of additional N stored in vegetation compared with the other drivers. Conversely, anthropogenic Nr substantially increases soil organic N—partly by decreasing the soil C : N—contributing thereby the largest share of the significant increase in soil N, whereas the effects of climatic, atmospheric CO₂ and land-use changes on soil N storage largely cancel out.

The average 2001–2010 rate of terrestrial nitrogen sequestration (27 Tg N yr⁻¹, figure 3) is a very small fraction of annual global terrestrial nitrogen turnover (about 800 Tg N yr⁻¹). This estimate is somewhat smaller than the 60 Tg N yr⁻¹ estimate by Galloway et al. [3] for the 1990s. However, Galloway did not separate sequestration from N exports owing to land use and land cover change (15 Tg N yr⁻¹), which compares well with the export number simulated by the ISAM-CN terrestrial biosphere model (15.6 Tg N yr⁻¹; [19]). The contribution of global N₉ deposition to the 2001–2010 N sequestration is 10 Tg N yr⁻¹ (figure 3), which is very close to the estimate of 9 Tg N yr⁻¹ by Schlesinger [20], based on the assumption that 50 per cent of the deposited N over forests would be sequestered. In O–CN, there is a large spatial gradient with close to 100 per cent retention in nutrient poor boreal systems and nearly no retention in nitrogen-saturated tropical and temperate ecosystems. Nitrogen retention is estimated to have declined globally from about 50 per cent in 1860 to 30 per cent at present. The estimated retention rate peaked in the 1980s with about 16 Tg N yr⁻¹, and remained high until the early 1990s when N deposition started to decline regionally (e.g. in Central Europe), and highly polluted ecosystems reached saturation. O–CN predicts gradually increasing terrestrial N losses since the 1950s, stagnating at year 2000 levels as a consequence of the estimate of declining global nitrogen deposition in 2001–2010 simulated by NCAR-CTM. This is a significant difference to cropland ecosystems, which show a modest and stable N sequestration rate since the 1980s, but strongly increasing leaching and volatilization losses with increasing fertilizer consumption.

(ii) Nitrogen–carbon couplings
The spatial pattern of N availability shows a strong latitudinal gradient, which is regionally dominated by the signature of the human N₉ perturbation due to deposition and fertilizers. Figure 4 displays the resulting patterns of contemporary N limitation of vegetation growth and carbon storage, which follows closely the pattern of N availability: the naturally high N limitation in the boreal and temperate zone due to low natural N fixation is regionally dominated by anthropogenic N₉ inputs. This regional pattern is consistent with current understanding [2,11], but difficult to evaluate quantitatively because of the lack of suitable observations. Hyper-spectral remote sensing might be one way forward, as it provides a direct measure of...
chlorophyll. However, a range of complicating factors in interpreting these data hamper their application at present [21].

Nitrogen additions to the terrestrial biosphere have increased global productivity by an estimated 2.6 Pg C yr\(^{-1}\), which corresponds to 2 per cent of the global annual total production and 12 per cent of the increase since pre-industrial times (table 2). About 0.2 Pg C yr\(^{-1}\) of this increased production is sequestered in the terrestrial biosphere, corresponding to 10–20 per cent of the global net land carbon uptake (table 2). Earlier studies based on simple biogeochemical models and upscaling of field-based carbon-sequestration estimates have estimated C sequestration based on N deposition estimates as 0.4–0.7 Pg C yr\(^{-1}\) in 1990 [5,22]. The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]). The estimate of the process-based O–CN model applied here is somewhat lower, but within the range of model simulations with the current generation of carbon–nitrogen cycle models (0.2–0.6 Tg N yr\(^{-1}\); [7]).

The additional carbon sequestration due to anthropogenic N\(_2\) additions has a perceivable but small cooling effect for the climate system, as it reduces the rate of atmospheric CO\(_2\) accumulation due to fossil-fuel burning. The nitrogen–carbon cycle interactions have further climate-relevant consequences, as increased plant N uptake due to CO\(_2\) fertilization reduces nitrogen losses globally, including the terrestrial N\(_2\)O emissions from soils (table 2). This counteracts the strong simulated increase in terrestrial N\(_2\)O emissions due to recent climate changes (0.8 Tg N yr\(^{-1}\); corresponding to an increase of 13% relative to pre-industrial conditions). Warmer temperatures will enhance nitrogen cycling and probably also N\(_2\)O production where N is not limiting [23]. However, there is mixed empirical evidence from ecosystem warming experiments, which show varying responses of soil N\(_2\)O emissions, resulting from the concurrent effects of changes in the moisture regime, plant and microbial N demand and biodiversity [24–27]. In agreement with earlier studies [5,6], the dominant cause for the estimated increase in terrestrial N\(_2\)O emissions is anthropogenic N\(_2\) inputs (table 2), which reduce or even overcompensate the climatic benefits from carbon sequestration in response to anthropogenic N\(_2\) inputs. There have also been slight increases in NO\(_x\) emissions from natural and fertilized soils (table 2), with as yet unquantified effects on the climate system. However, this anthropogenic soil NO\(_x\) source remains small compared with
Table 2. Attribution of the changes in the global nitrogen budget from 1860 to 2010 due to changes in land cover and land-use (LUCC), increased atmospheric CO₂ abundance (CO₂), climatic variability and changes (climate), anthropogenic reactive nitrogen additions (deposition) and industrial fertilizer application (fertilizer). Note that this analysis does not take account of manure additions. Land-use emissions in 2000–2010 are an underestimate, as the dataset for land-use changes stops in 2005 [13]. Values are reported from simulations driven with the NCAR (TM5) nitrogen deposition fields. GPP, Gross primary production; NBP, net biome production = net ecosystem production – anthropogenic C losses.

<table>
<thead>
<tr>
<th></th>
<th>1860s</th>
<th>2000s</th>
<th>LUCC</th>
<th>CO₂</th>
<th>climate</th>
<th>deposition</th>
<th>fertilizer</th>
</tr>
</thead>
<tbody>
<tr>
<td>fluxes</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>natural N input</td>
<td>112.4 (112.7)</td>
<td>98.7 (99.0)</td>
<td>−10.0</td>
<td>3.1</td>
<td>0.1</td>
<td>2.5 (−2.3)</td>
<td>0.1</td>
</tr>
<tr>
<td>anthropogenic N</td>
<td>28.5 (22.5)</td>
<td>161.6 (155.2)</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>42.1 (44.0)</td>
<td>85.6</td>
</tr>
<tr>
<td>ecosystem N lossesa</td>
<td>148.2 (144.4)</td>
<td>237.1 (231.6)</td>
<td>19.2</td>
<td>−18.9</td>
<td>11.4</td>
<td>29.9 (31.2)</td>
<td>50.0</td>
</tr>
<tr>
<td>soil NO₃ emission</td>
<td>6.5 (6.3)</td>
<td>8.5 (8.4)</td>
<td>0.1</td>
<td>0.2</td>
<td>0.8</td>
<td>0.7 (0.7)</td>
<td>0.7</td>
</tr>
<tr>
<td>soil N₂O emission</td>
<td>6.0 (5.8)</td>
<td>8.7 (8.6)</td>
<td>0.3</td>
<td>−0.6</td>
<td>0.8</td>
<td>0.8 (0.8)</td>
<td>1.5</td>
</tr>
<tr>
<td>GPP</td>
<td>99.9 (100.0)</td>
<td>122.3 (121.6)</td>
<td>−0.4</td>
<td>17.1</td>
<td>2.9</td>
<td>1.66 (1.81)</td>
<td>0.93</td>
</tr>
<tr>
<td>NBP 1990s</td>
<td>1.3 (1.3)</td>
<td>−0.8</td>
<td>2.0</td>
<td>0.0</td>
<td>0.20 (0.21)</td>
<td>0.07</td>
<td></td>
</tr>
<tr>
<td>NBP 2000s</td>
<td>1.9 (1.9)</td>
<td>−0.4</td>
<td>2.4</td>
<td>−0.2</td>
<td>0.18 (0.17)</td>
<td>0.06</td>
<td></td>
</tr>
<tr>
<td>pools</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>vegetation N</td>
<td>3.4 (3.6)</td>
<td>3.4 (3.4)</td>
<td>−0.71</td>
<td>0.36</td>
<td>0.04</td>
<td>0.09 (0.10)</td>
<td>0.03</td>
</tr>
<tr>
<td>vegetation C</td>
<td>615.7 (607.2)</td>
<td>581.4 (578.1)</td>
<td>−141.3</td>
<td>98.6</td>
<td>3.8</td>
<td>2.7 (2.7)</td>
<td>0.08</td>
</tr>
<tr>
<td>soils N</td>
<td>103.2 (101.5)</td>
<td>103.9 (102.2)</td>
<td>−0.50</td>
<td>0.83</td>
<td>−0.39</td>
<td>0.61 (0.60)</td>
<td>0.56</td>
</tr>
<tr>
<td>litter and soil C</td>
<td>1312.3 (1291.3)</td>
<td>1287.6 (1268.2)</td>
<td>−52.1</td>
<td>35.9</td>
<td>−9.1</td>
<td>6.1 (5.8)</td>
<td>2.4</td>
</tr>
</tbody>
</table>

*Leaching + volatilization (NH₃ + NOₓ + N₂O + N₂).*

the anthropogenic NOₓ from combustion sources, which globally has a strongly negative effect on the climate forcing [28]. While these changes all matter to the climate system, the net effect of anthropogenic N deposition on climate system is still unknown [29].

(b) Future projections of coupled nitrogen–carbon cycle dynamics

Figure 5 illustrates the development of nitrogen limitation on terrestrial plant production and carbon sequestration between 1950 and 2100, based on projections with the SRES A’ scenario. Increasing atmospheric CO₂ enhances plant productivity and, therefore, N demand, which increases nitrogen limitation, as the higher demand cannot be completely met by reduced nitrogen losses, increasing N deposition or biological N fixation. This additional limitation is most pronounced in the boreal zone, where N constraints attenuate the direct CO₂ fertilization effect on plant production by more than 50 per cent and on carbon sequestration by nearly 80 per cent (in the year 2100) relative to a projection not taking N limitation explicitly into account (figure 5c,f). These projections are broadly consistent with the importance of a nitrogen constraint in free-air CO₂ enrichment experiments [30,31], and the simulated geographical distribution of nitrogen limitation (figure 4).

Consistent with observational evidence from a temperate forest soil warming study [32], warming increases carbon sequestration because of the remineralization of nitrogen from soil, which fertilizes the vegetation and thus increases accumulation of biomass. This climate effect is simulated to have increased productivity during most of the twenty-first century in the boreal and temperate zones, but the global effect is rather small because of opposite trends in tropical regions related to increased respiration costs. The same processes operate in two other global modelling studies [33,34]. However, these two studies suggest a stronger positive effect from climate change, such that in these studies the total carbon balance in the year 2100 is changed from a negative carbon balance to a positive carbon balance owing to the considerations of nitrogen–carbon cycle interactions.

Nitrogen deposition is estimated here to play only a small role in future carbon uptake (figure 5), as also reported by a simulation with the CLM4 terrestrial biosphere model [33]. The C sequestration resulting from anthropogenic N deposition (27 Pg C, sequestering also 3.9 Pg N) is slightly less than the sequestration resulting from climate change induced remineralization of soil N and enhanced vegetation growth (44 Pg C, recapturing 1.2 Pg N). A thorough analysis of the effects of future N deposition is still lacking. However, given these results, N deposition needs to be a component of future global carbon-cycle projections.

N limitation of CO₂ fertilization dominates the estimated long-term trend of terrestrial carbon sequestration at all latitudinal bands (figure 5), consistent with two other independent modelling studies [33,34]. The dominance of the reduced CO₂ fertilization due to N limitation has important consequences for projections of future climate changes with interactive biogeochemical cycling: neglecting an explicit treatment of N dynamics in coupled carbon-cycle climate modelling studies such as the Coupled Climate–Carbon Cycle Model Intercomparison Project (C4MIP) [35] will lead to an underestimation of the build-up of fossil CO₂ in the atmosphere [7]. For the O–CN model and the SRES A2 scenario, nitrogen dynamics reduce global carbon sequestration between 1860 and 2100 by 164 Pg C (358 Pg C for the CO₂ fertilization effect only), because of a regional nitrogen deficit of
5.7 (12.0) Pg N. Depending on whether the radiative forcing in Earth system models is prescribed (RCP-type forcing) or calculated based on the greenhouse gas and aerosol burden of the atmosphere, neglecting nitrogen–carbon cycle interactions will lead to an underestimation of the need for emission reductions of carbon-sequestration efforts to meet a certain radiative forcing pathway, or the rate of climate change, respectively. Remineralization of nitrogen due to accelerated soil organic matter turnover, and deposition of reactive nitrogen deposition are not strong enough to counteract this phenomenon, even though they lead to increased carbon sequestration.

The future changes in the terrestrial nitrogen and carbon balance also induce changes in NO
\(_x\) and N\(_2\)O emissions from soils. O–CN suggests a change of -3.1 (-0.8) Tg N yr\(^{-1}\) from pre-industrial to 2100 soil N\(_2\)O emissions due to climate change (CO\(_2\) fertilization), with similar changes occurring also for the terrestrial soil NO\(_x\) source. This results implies a positive terrestrial N\(_2\)O-climate feedback of 0.54 Tg N yr\(^{-1}\) K\(^{-1}\), which would be weakened by a smaller negative carbon-concentration-N\(_2\)O feedback. However, one should place limited confidence in this estimate from one model and one scenario. A feedback of this magnitude would be important enough to require further consideration in coupled biogeochemistry–climate models, even though the biospheric feedback might, as with anthropogenic CO\(_2\), be small compared with future anthropogenic emissions of N\(_2\)O from managed ecosystems [36].

4. Discussion

This study provides an advance over previous assessments [3,20], as it relies on a process-based ecosystem model that integrates the key carbon–nitrogen cycle interactions and their coupling to biogeophysical processes, while considering the impacts of atmospheric (climate, CO\(_2\)) and land cover changes. Tables 1 and 2 provide an assessment of the uncertainties related to estimates in nitrogen deposition, and show that the simulated trends and spatial patterns are reasonably robust against these uncertainties. Regionally important ecosystem types (e.g. wetland and peatland ecosystems [37]), land management characteristics (such as nitrogen efficient farming, manure-based agriculture [38]) and effects of...
N-related air pollution (such as tropospheric ozone [39]) have been neglected, because they cannot be simulated by the current version of O–CN, but they might nonetheless be globally significant.

The increased complexity of the analyses introduces new uncertainties. While the simulated trends are considered robust, other carbon–nitrogen cycle models may give notably differing estimates. Key uncertainties in the modelling include: (i) the response of canopy-level photosynthesis to nitrogen additions; (ii) changes in the allocation patterns (root:shoot ratio); (iii) the competition of plants and soil microbes for the added (or reduced) amount of nitrogen, and therefore the temporal dynamics of the fate of the added N; (iv) the change of ecosystem stoichiometry over time; (v) responses of and controls on biological N fixation; and (iv) the fraction of N that is exported from ecosystems. Evaluation against ecosystem manipulation experiments, which were part of the O–CN model evaluation [8,10], help to understand whether the model’s sensitivities to perturbations are adequate. However, the interpretation of these experiments is complex, and their regional representativity unclear, such that, whereas O–CN’s sensitivities appear reasonable, large uncertainties remain in the modelled responses, requiring further assessments.

Another factor omitted in this assessment is the co-limitation of the terrestrial nitrogen and carbon cycles by phosphorus. Plants have evolved strategies to access soil P using phosphatase exudation, such that P limitation mainly occurs on old, deeply weathered and P-deprived soils [40]. The results presented here are consistent with the hypothesis [41,42] that temperate and boreal ecosystems are limited by N, whereas moist tropics are not. Given that most of the anthropogenic perturbation of the nitrogen cycle so far occurred in predominantly N-limited regions, it is unlikely that the analyses of the fate of anthropogenic N and its consequences for the carbon cycle would be dramatically altered when accounting for the phosphorus cycle. However, future projections of the global carbon cycle will be different in regions where P limitation prevails.

## 5. Concluding remarks

The estimates presented in this study result from a state-of-the-art terrestrial biosphere model integrating biophysical, biogeochemical and ecological process understanding. There is considerable uncertainty in any such model and a systematic assessment of nitrogen–carbon cycle interactions by an ensemble of such models seems the logical next step to take. Nonetheless, some conclusions appear robust:

- anthropogenic N additions currently enhance nitrogen and carbon sequestration in the biosphere (figure 1), but cause at the same time increased emissions of NOx and N2O from soils. Each of the factors is large enough to matter to the climate system, but the net climatic effect is still uncertain;
- nitrogen is limiting terrestrial productivity in many ecosystems, and therefore the capacity of the terrestrial biosphere to sequester carbon in response to increased atmospheric abundance of CO2;
- regional and global strategies for increasing terrestrial carbon storage in either woody biomass or soils need to consider the consequences for nutrient cycling and anticipate the effects of nutrient limitation when discussing the effectiveness of different measures; and
- future projections of the global carbon cycle will underestimate the fraction of anthropogenic fossil-fuel-based CO2 emissions remaining in the atmosphere, unless nitrogen dynamics are taken into account. Because of the tight coupling of the terrestrial nitrogen and carbon cycles and their interactions with climate, nitrogen dynamics need to be accounted for interactively in the next generation of Earth system models designed for long-term studies of biogeochemical–climate interactions.

S.Z. was supported by the Marie Curie Reintegration Grant JULIA (PERG02-GA-2007-224775) and the European Community’s Seventh Framework Programme under the GREENCYLCES II ITN (grant agreement no. 238066) and the ECLAIRE project (grant agreement no. 262910).

## References

35. Friedlingstein P et al. 2006 Climate-carbon cycle feedback analysis: Results from the CCMIP model intercomparison. J. Clim. 19, 3337 – 3353. (doi:10.1175/1175/JCLI3880.1)